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Iryna Matveeva

OCAPACITY MODEL APPLIANCE FOR ATION OF RADIATION FACTOR IN A LAKE ECOSYSTEM

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servich to creation of an ecological standardization system is offered. Such system services is and ardization system. Basing on the radiation capacity model and theory sefining of critical biota condition on an example of a lake ecosystem is offered. If a critical radiation dose on the given biota, through radiocapacity models the pollution of the ecosystem's components are determined. The levels of permissible seed as such, at which radiation dose on the critical biota of the lake may not exceed

Landion doses on the biota; radiocapacity of ecosystems.

standardization of radionuclides release tant, because there is a masse processes not only ach dominate in modern cology concerning for the biota in case of observing not always true, is not always twed by modern researches and all it is connected with the facts avoid negative influence of the usually cannot do this.

__i publications

rical paradigm, when the safety of is defined by norms of people's ploitation, doesn't stand up to why the International Comission Protection in publication №80 set a system of ecological based on reaction of the biota on Lation doses. [2, 4]. The variants are in this article, when the radiological rdization on the biota is estimated in eaction of the critical (the most (i.e.) kinds of organisms. But it is to be the critical kind for one ecosystem may atical for other types of ecosystems [2, 4]. it became necessary to develop other es. The substantiation of this need follows : this article.

The objective of the work: to built the approaches to principles of substantiation and creation of ecological norms on permissible levels of radiative effect on the ecosystems' biota, which differs from the current hygienic rating system. The thing is that in the situations of hygienic standards adherence the dose effects on the ecosystems' biota may be critical.

3. Results of the research

In the most radiological situations the biota in the environment where it grows is exposed to external (from the exposure sources outside the biota) and internal irradiation (from incorporated in the tissues radionuclides). In the irradiation biocenosis for the experimental organisms the exposure sources may become incorporated (accumulated) radionuclides, contained in contiguous organisms. For certain organs of plants and animals external are also those sources, which are contained in other parts of this plant or animal.

In case of biocenosis contamination by artificial radionuclides initially the radioactive substances are contained on the ground and water surface contacting with plants or animals. Only after a certain time period under influence of wind, downfalls, biomass increase radionuclides are being reallocated within the abiotic constituent of the ecosystem and as result of migration processes or anthropogenic measures shift into the depth of the ground and body of water.

In case of radionuclide emissions in the environment the need arises to define the boundary values of radionuclides intake into the ecosystem,

sed significant changes in the

e waluating the maximum release into the ecosystem

annual exposure rate. In the mov and V.G. Tsytsuginoy [2] on ecosystems in the form introduced (Table 1). s that the actual dose limit for e and "storage" in ecosystems may be dose rate that does not **★** vear when visible ecological **red** on the scale. According to the coround exposure rate of 0.4-4.0 to the concentration of ¹³⁷Cs **k**Bq kg in the ecosystem or its and aquatic plants) and about e ecosystem including terrestrials, Bg/kg in average. Calculations s of dose factors developed by ented in Table 2 [1].

Table 1. Scale of radiation doses and zones in ecosystems

Number of dose limit	Zone	Exposure rate, Gy/year
1	Zone of radiation safety	< 0,001 — 0,005
2	Zone of physiological masking	> 0,005 0,05
3	Zone of ecological	
	masking:	
3.1	terrestrials	> 0,05 0,4
3.2	hydrobionts and	> 0,05 4
	terrestrials	
4	Zone of visible	
	ecological effects:	
4.1	dramatic for terrestrials	≥ 0,4
4.2	dramatic for	≥ 4
	hydrobionts and	
	terrestrials	
4.3	disastrous for animals	≥ 100
	and plants	

Table 2. Dose factors for the biota ecosystems on some radionuclides

	Internal irradiation,	External irradiation				
	Gy/year/Bq/kg	Water,	Air,	Ground,	Vegetation,	
		Gy/year/Bq/m ³	Gy/year/Bq/m ³	Gy/year/Bq/kg	Gy/year/Bq/kg	
	4,1 10-6	2,7 10-9	1,72 10 ⁻⁶	4,02 10 ⁻⁶	1,72 10 ⁻⁶	
. H	2, 88 10 ⁻⁸	0	0	0	0	
ales. Mary	3,44 10-6	1,76 10 ⁻⁹	1,43 10 ⁻⁶	2,64 10 ⁻⁶	1,43 10 ⁻⁶	
-	3,52 10 ⁻⁶	1,57 10 ⁻⁹	1,43 10 ⁻⁶	2,36 10 ⁻⁶	1,43 10 ⁻⁶	
	2,86 10 ⁻⁵	1,48 10 ⁻¹⁰	7,73 10 ⁻⁸	2,22 10 ⁻⁷	7,73 10 ⁻⁸	
	2,64 10 ⁻⁵	3,72 10 ⁻¹²	2,35 10-9	5,58 10 ⁻⁹	2,35 10 ⁻⁹	
	9,92 10 ⁻⁷	3,07 10 ⁻¹⁰	2,83 10 ⁻⁷	4,61 10 ⁻⁷	2,83 10 ⁻⁷	
÷.:	1,12 10 ⁻⁴	8,91 10 ⁻⁹	6 10 ⁻⁶	1,43 10 ⁻⁵	6 10 ⁻⁶	
-, -	2,5 10 ⁻⁷	6,51 10 ⁻¹²	6,01 10 ⁻⁹	9,77 10 ⁻⁹	6,01 10 ⁻⁹	

Let data in Table 2 allow us to calculate the Let doses of the wild biota in different types of Let doses.

standardization in the lake stem. The simulation results of permissible in lake ecosystem [3]

According to the estimation of the maximum massible radionuclides ¹³⁷Cs concentration in the ments of the ecosystem the critical releases and massions into ecosystems can be estimated (we will with an example — the lake). Based on the model adiocapacity of the lake ecosystem [3] we showed for benthos of benthal deposits of a freshwater

basin maximum permissible radionuclides release into the pond (N_k) shall not exceed the following value:

$$N_k < \frac{LHS}{kF}$$
,

where L — calculated based on the dose limit of 4 Gy/year (limit) of 137 Cs radionuclides concentration in aquatic biota (600 kBq/kg);

S — area of the basin;

H— the average depth of the basin;

k — coefficient of radionuclides accumulation from water by benthal deposits;

F — radiocapacity factor of benthal deposits of the basin.

* **the water** column the

Cs accumulation $S = 2 \text{ km}^2$, where $S = 2 \text{ km}^2$, S = 0.7 critical value of to the calculations than S = 0.7 TBq into

a critical value of basin for its benthos is

less than the permissible er of this lake, which is of the water column of the

we give a specific example nodel to the lake ecosystem. Elculation of permissible not the lake.

into the lake area of 1 km² MBq of ¹³⁷Cs.

5 m. the thickness of the layer

10 cm, the sludge

(AC) — 200 and into account

10 AC of the biota, living in the

10 200 and 1 to 100000.

what amount of radionuclides can the lake, so that the dose for the not exceed the critical limit AC

above mentioned formula we will Sle ¹³⁷Cs releases (Table 3).

in was performed as follows.

consistent pattern of radionuclides rough the components of the lake can define the radioactivity levels in

Table 2), we can calculate the partial biota from different components of the stem for different values of AC for the bota.

at the value of the initial radionuclides only 1 MBq ¹³⁷ Cs.

Then we take, for example, the summary dose in the last column, which is equal to $4.7 \cdot 10^{-3}$ Gy/year when the release is 1 MBq.

And if the permissible dose to the seabed biota, as we have defined, should not exceed 4 Gy/year, then dividing the value of 4 Gy/year by the estimated value of $4.7 \cdot 10^{-3}$ Gy/year we become the estimation of the quantity — Bq in permissible release, which makes — $8.5 \cdot 108$ Bq/year or 0.023 Cu/year.

Thus, it should be noted that at very high values of AC of the biota (100 000 units) annual permissible radionuclides release in the researched lake can make a very small value, only 0,023 Cu/year on only 1km² area of the lake.

A similar calculation was made for another biogenous radionuclide — 90 Sr.

It can be seen that, depending on the AC of the biota permissible releases into this lake make for ¹³⁷Cs from 0,023 to 2100 Cu per year, and for ⁹⁰Sr — from 0.1 to 7800 Cu per year, if releases take place for only one year, as it was after the Chernobyl disaster.

If it is a working nuclear power plant, it is clear that the permissible releases per year will be significantly less, so that they do not lead to exceeding the dose limits.

That is, for real values of AC for the seabed biota fairly strict limits on permissible levels of releases into such lake ecosystem can be applied.

In most cases the levels of water pollution, for which there are hygienic standards (2 Bq/L for ¹³⁷Cs), will be estimated as significantly smaller than the given hygienic standards.

Thus, the analysis shows that actually in this case of the lake ecosystem the ecological norm can be estimated as much tighter than the known hygienic standard.

4. Conclusions

Generally in ecology and radioecology in ecological standardization, environmental dominates such paradigm. If for human it is comfortable to live in the ecological situation, all the more nothing will damage the biota.

The analysis, which we conducted shows that this is not really so.

That is a safe situation for human can turn into high doses for the biota due to redistribution of radionuclides and high values of AC, which are peculiar to biota.

Table 3. Calculation of the dose on the components of the lake ecosystem and repermissible annual ¹³⁷Cs release depending on the values of AC for the benthos biota

enents of the lake	K_H of the biota of benthal deposits of the lake (benthos)					
	1	10	100	1000	10000	100000
	5,4–9	5,4 –9	5,4-9	5,4-9	5,4–9	5,4-9
	3,2–8	3,2-8	3,2–8	3,2–8	3,2–8	3,2–8
s of the lake	1,4-8	1,4–7	1,4-6	1,4–5	1,4-4	1,4–3
	3,3-8	3,3–7	3,3-6	3,35	3,3-4	3,3-3
e biota	5,2-8	4,8–7	4,7–6	4,7–5	4,7–4	4,7-3

ease into the lake:

7,7+13 Bq 2100 Cu	8,4+12 Bq 220 Cu	8,4+11 Bq 22 Cu	8,5+10 Bq 2,3 Cu	8,5+9 Bq 0,23 Cu	8,5+8 Bq 0,023 Cu
2,9+14 Bq 7800 Cu	3,8+13 Bq 1020 Cu	3,9+12 Bq 105 Cu	3,9+11 Bq 10,5 Cu	3,9+10 Bq 1 Cu	3,9+9 Bq 0,1 Cu

the lake when hygienic standards for an be easily fulfilled and the limits the biota of the lake can be should be emphasized that the se limits on the biota of benthal to death of a part of the biota, and will lead to acidification of the (pH may drop to values 5-6), turn can cause desorption of cumulated in the benthal deposits.

The may also obviously exceed hygienic

that the establishment of actually gical standards for Ukraine and other ery complicated task.

- .m is that it is nearly impossible to .d ecological standards for permissible releases in different ecosystems.
- e. generally any ecosystem will require nent of a specific model and evaluation at value of ecological standard.

roblem remains, and it is necessary to

These same problems arise for other types of ecosystems as well.

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e. Proceedings of the National Aviation University. 2014. N 1 (58): 70–74 рісмності для нормування радіаційного фактора в озерній 🖚 Космонавта Комарова, 1, Київ, Україна, 03680 стеми екологічного нормування. Показано, що така система не радіаційної ємності і теорію установлення критичного стану біоти на прикладі озерної п-ного дозового навантаження на біоту досліджуваної екосистеми, топустимого забруднення радіонуклідами компонентів екосистеми. дення оцінено як такі, при яких дозове навантаження на критичну то боту: екологічний норматив; радіоємність екосистем. **ръ**ли радиоемкости для нормирования радиационного фактора в **штет. пр**осп. Космонавта Комарова, 1, Киев, Украина, 03680 р системы экологического нормирования. Показано, что такая система не тигненического нормирования. Опираясь на модели радиационной емкости 🗷 📭 сложен метод установления критического состояния биоты на примере 📭 н контроль критической дозовой нагрузки на исследуемую биоту, через **ур**овни допустимого загрязнения радионуклидами компонентов экосистемы. загрязнения оценены как такие, при которых дозовая нагрузка на тревышать предельно допустимую норму — 4 Гр/год. жин на биоту; радиоемкость экосистем; экологический норматив. e of Engineering, Associate Professor. Aviation University, Kyiv, Ukraine. ogical Institute, Ternopil, Ukraine (1990). **and application** of the radiation capacity model for normalization of radiation factor in